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Metal-Fungus interaction: Review on cellular processes underlying heavy metal detoxification and synthesis of metal nanoparticles

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- 1 Metal-Fungus Interaction: Review on cellular processes underlying heavy metal
- 2 detoxification and synthesis of metal nanoparticles
- 3
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14 Abstract

15 The most adverse outcome of increasing industrialization is contamination of the ecosystem with heavy metals. Toxic heavy metals possess a deleterious effect on all forms of biota; however, 16 they affect the microbial system directly. These heavy metals form complexes with the microbial 17 system by forming covalent and ionic bonds and affecting them at the cellular level and 18 biochemical and molecular levels, ultimately leading to mutation affecting the microbial 19 population. Microbes, in turn, have developed efficient resistance mechanisms to cope with 20 metal toxicity. This review focuses on the vital tolerance mechanisms employed by the fungus to 21 resist the toxicity caused by heavy metals. The tolerance mechanisms have been basically 22 23 categorized into biosorption, bioaccumulation, biotransformation, and efflux of metal ions. The mechanisms of tolerance to some toxic metals as copper, arsenic, zinc, cadmium, and nickel have 24 25 been discussed. The article summarizes and provides a detailed illustration of the tolerance 26 means with specific examples in each case. Exposure of metals to fungal cells leads to a response

that may lead to the formation of metal nanoparticles to overcome the toxicity by immobilization in less toxic forms. Therefore, fungal-mediated green synthesis of metal nanoparticles, their mechanism of synthesis, and applications have also been discussed. An understanding of how fungus resists metal toxicity can provide insights into the development of adaption techniques and methodologies for detoxification and removal of metals from the environment.

Keywords: Fungus, Metal resistance, Heavy metals, Metal adsorption, Bioaccumulation, Metal
 nanoparticles

34

35 **1. Introduction**

Heavy metal pollution is a significant concern in many countries where many industrial effluents 36 37 comprise toxic metal ions. The extent of absorption and distribution, species of origin, bioavailability, ubiquity, and degree of use determines the toxicity of a metal (Zaidi and Pal, 38 2017). The highly persistent nature of these pollutants results in increased accumulation in soil 39 and water bodies (Zafar et al., 2006), making bioremediation difficult. As per the World Health 40 Organization (WHO), metals such as cadmium (Cd), lead (Pb), nickel (Ni), cobalt (Co), and 41 copper (Cu) exhibit severe toxicological effects on aquatic animals and microorganisms. 42 Therefore, the WHO has stipulated tolerable limits of these heavy metals in drinking water. Such 43 maximum permissible limit for specific heavy metals are as follows, Cu 2.0 mg/l; Mercury (Hg) 44 45 0.006 mg/l; Cd 0.003 mg/l; Arsenic (As) 0.01 mg/l; Pb 0.01mg/l; Chromium (Cr) 0.05 mg/l and Nickel (Ni) 0.07 mg/l (WHO, 2011). However, innate defense strategies of prokaryotes allow the 46 uptake of heavy metal ions and their utilization as nutrients by absorptive and accumulative 47 capacity. Likewise, a eukaryotic microbial system such as fungus has been reported to have 48 enough strategies to survive with metal stress. The different mechanism utilized by the fungus to 49

50 survive in the presence of metals includes binding of metal ions to high-affinity functional 51 groups, metal accumulation, complexation with different microbial biomolecules and active 52 efflux or exclusion of metals. This review focuses on the detailed investigation and comparative 53 analysis of the different mechanisms employed by the fungus to survive in a hostile metal-54 contaminated ecosystem. The toxicological impact of heavy metals and the subsequent strategies 55 used by fungal species to limit their effects are described in the subsequent sections with 56 examples of fungal species exhibiting resistant strategies.

57 **2.** Toxic effect of heavy metals on fungus

Both natural and anthropogenic activities can be articulated for the occurrence and accumulation 58 of metals in the environment. Effluents from mineral, tanning, automobile, fossil fuels, 59 fertilizers, agricultural, and petroleum industries are few anthropogenic means that significantly 60 contribute towards the release of metals in the atmosphere (Basta et al. 2004; Alloway et al. 61 1995). These effluents constituted of Cu, manganese (Mn), Hg, Pb, and zinc (Zn), along with 62 other inorganic components, which are not degradable and, hence, retained by the soil 63 64 (Kirpichikova et al. 2006). Because of the high persistence of these metal ions, they accumulate over time with soil acting as a reservoir or metal sink (Adriano, 2003). Industrial and 65 anthropogenic activities are major contributors of heavy metal accumulation in environment, 66 whose excessive presence in soil have adverse effect on fungal growth, and microbial-plant 67 68 symbiotic association thereby affecting soil fertility, uptake of minerals and nutrient availability to plants (Miransari 2016). 69

70 Variation in soil physicochemical parameters such as pH and redox potential leads to
71 immobilization of metal ions that exert toxic effect on soil microbes (Gupta 1992). The

interaction of fungus with toxic heavy metals induces stress conditions leading to a significant 72 reduction in microbial growth, thereby exerting selective pressure on the microbiota (Nwuche 73 and Ugoji, 2008). Excessive concentration of accumulated metals causes microbial membrane 74 damage, organelle damage, lipid peroxidation and cell apoptosis contributed by generation of 75 reactive oxygen species (Igiri et al. 2018). Literature also signifies that excessive concentration 76 77 heavy metal can inhibit mycorrhizal fungus functioning (Del et al, 1999). Additionally, pH also affects availability of metals and microbial population in soil. Alkaline pH is known to restrict 78 79 microbial growth and subsequently decreases the amount of metals (Cd and Zn) available to plants by action of symbiotic fungus (Wang et al, 2006), due to immobilization of metal ions due 80 to precipitation. Soil pH is also known to affect the availability of certain heavy metals ions. A 81 positive correlation between alkaline pH and metal concentration above allowable limit has been 82 suggested (Lenart and Wolny-Koladka, 2012). 83

The fungus is known to utilize both metabolism dependent and independent means to survive under high metal stress conditions (Akar et al., 2005). Table 1 summarizes the various metalbinding sites in fungal cells.

As evident the accumulated metal ions introduce several physiological and morphological changes in the microbes. Similarly, toxicity of metal ions is observed at both cellular and molecular levels. Broadly, the cellular defense mechanisms exhibited by fungus can be categorized into extracellular and intracellular means. While the former inhibits the uptake and internalization of metal ions, the later reduces the toxic effect of entrapped metals by binding with biomolecules or efflux channels.

94 **2.1. Effect on fungal cell growth**

Studies have reported that heavy metal ions adversely affect the fungal cellular life process, 95 ultimately leading to growth inhibition and death. The growth of various species of Aspergillus 96 fungus and Sterigmatomyces halophillus were observed to be inhibited by heavy metals like Cu, 97 Cd, Mn, Pb, and Zn (Bano et al. 2018). Pb inhibits the growth of Aspergillus biennis at a 98 concentration of 10 to 30 mM range (Graz et al. 2011), while Cd inhibits the growth of 99 100 Schizophyllum commune at a concentration of 0.1–0.2 mM (Lilly et al. 1992). Ni induces pyruvate release by altering carbohydrate metabolism, thereby disintegrating the membrane 101 structure in Neurospora crasa (Mohan and Sastry 1984). Excessive accumulation of Cd in 102 103 microbes induces protein and nucleic acid damage resulting in inhibition of cell division, as well as, transcription. At the same time, Hg ions are known to disrupt the microbial cell membrane, 104 impair enzyme function, and denature proteins (Ayangbenro and Babalola, 2017). 105

106 **2.2.** The effect at the biochemical and molecular level

107 Studies report that a high concentration of toxic metals inhibit enzyme activity, alters enzyme specificity, and leads to conformational changes in nucleic acid and proteins affecting transcriptional and 108 109 translational steps (Bong et al. 2010). Furthermore, metal ions form precipitates or complexes with 110 microbial metabolites leading to disruption of cellular organelles and whole-cell membranes (Sobolev and Begonia 2008). Ni is reported to induce the deregulation of iron and organic acid 111 metabolism in A. niger (Yang et al. 2016; Adiga et al. 1961). Elevated concentration of Cu 112 113 disintegrates membrane integrity causing the release of solute molecules from cells, ultimately inducing cell death. Cu is also known to cross-link between the DNA strands distorting the 114 helical structure of DNA (Rifkind et al. 2001). Griffiths et al. (1997) assayed the % (G+C) 115 profiling and community hybridization by extracting DNA from a contaminated abatable soil 116

sample and suggested that Cd contamination brings about a decrease in carbon biomass. Cd at a 117 concentration of 50 mg/kg diminishes enzyme activity, precisely that of dehydrogenase, alkaline, 118 and acid phosphatase in sludge (Dar 1996; Landi et al., 2000). A schematic representation of the 119 mode of action of heavy metals on fungus has been illustrated in Figure 1. Exposure of fungal 120 species to heavy metal ions thus induces stress conditions resulting in drastic increase in 121 122 production of reaction oxygen species as superoxide, peroxides and hydroxyl radicals, that damage fungus cell, organelle structure and alter metabolism (Luna et al 2015, Zang et al 2015, 123 124 Lin et al 2020). Thus, oxidative stress can be considered as prime reason behind heavy metal 125 induced cell death in microbes.

126

127 **3. Fungal tolerance to heavy metals**

As heavy metals are potentially toxic to fungus leading to cell lysis and death, numerous studies 128 focused on deducing the extent of tolerance of various fungi. It was inferred that the specificity 129 130 and maximum tolerance limit for metals greatly vary at the genus and species levels. Fungus belonging to the species Aspergillus sp. and Penicillium sp. are among the most resistant strains 131 (Ezzouhri et al. 2009). The difference in the tolerance level of fungal strains (Aspergillus sp, 132 *Penicillium* sp, and *Fusarium* sp) isolated from the same location was due to the different 133 resistance strategies employed to tolerate metals (Jacob et al., 2013). Aspergillus flavus (ASC1) 134 135 and A. niger (ASB3) were capable of growing in As supplemented media up to a concentration of 2000 ppm and 1200 ppm, respectively (Mukherjee et al. 2013). While Aspergillus sp. was 136 137 resistant to Pb, Cr, Cu, and Zn, In another similar study, A. niger, A. flavus, and Fusarium sp. 138 were reported to be most tolerant to Cr and Pb (Iram et al. 2013).

Penicillium chrysogenum and Trichoderma viride were reported to tolerate Cr up to a concentration of 600 mg/l (Levinskaite, 2002). Few other strains of *Fusarium* sp. and *Penicillium* sp. can tolerate up to 100mg/l Cr while *A. flavus* tolerated up to 400 mg/l Cr (Iram et al. 2012). Similarly, a study comprising 36 strains of *Anthrodia vaillantii* showed that some were able to tolerate up to 40 mM Cu. In contrast, the growth of other strains was inhibited at a mere 3 mM concentration indicating variance in metal tolerance capacity (Collet, 1992).

145 Further, there have been disparities in the tolerance limit of fungal strains isolated from metal contaminated and non-contaminated sites (Howe et al. 1997). Suillus luteus fungi, isolated from 146 metal contaminated sites, were observed to be more tolerant than those isolated from non-147 contaminated sites (Colpaert et al. 2000). However, no relation was observed in a Cd resistant 148 fungus P. betulinus isolated from the contaminated and non-contaminated site (Baldrian and 149 Gabriel 2002). The above reports suggest that the striking difference in metal tolerance capacity 150 and specificity. This is probably due to the use of different strategies to resist the toxic effect of 151 metals. Additionally, physiochemical properties of soil and carbon content also affect the 152 153 capacity of metal tolerance in fungus. A high content of carbon and acidified soil favors fungal 154 growth and so fungal community growing under such conditions have increased metal tolerance. 155 Table 2 clearly shows the variance in the type of metal tolerated by different fungal strains and 156 also gives information on how well they tolerate a particular metal ion or range of metal ions.

4. Mechanism of heavy metal resistance in fungus

The development of metal resistance in microorganisms has evolved due to their continual exposure to toxic metals since life started at least 3.5 billion years ago. Interaction between metals and fungus depends on several factors such as type and concentration of metal, organism and the nature of the contaminated environment. In response to the presence of undesirable

metal, fungi have developed both intracellular and extracellular strategies to resist its toxic 162 effect. A detailed investigation of the heavy metal tolerance in fungi suggested the involvement 163 164 of extracellular system as metal chelation and cellular binding, wherein the entry of the metal ion to the cell interior is prevented. Intracellular strategies involved the conjugation of the metal ion 165 with fungal biomolecules like proteins and organic ligands. On the basis of this, we have broadly 166 167 categorized the defense mechanism used by fungus under four subheadings (1) biosorption (2) bioaccumulation and compartmentalization (3) metal chelation, and (4) efflux transport for metal 168 169 exclusion. Figure 2 depicts the different strategies employed by fungus for resistance to toxic 170 heavy metals.

171

4.1. Biosorption

Biosorption describes the adsorption of metal ions via physicochemical interaction with fungal 173 cellular compounds (Kapoor and Viraraghavan, 1998). Biosorption prevents the uptake of metal 174 175 ions into cell interior aided by the interaction of metals with the microbial surface through electrostatic interactions, Van der Waals forces, covalent bonding, or a combination of these. 176 The microbial cell surface has a net negative charge due to the presence of biomolecules 177 polysaccharides, proteins, and phenolic acids, which have functional groups like hydroxyl, 178 179 carbonyl, amine and phosphoryl groups, thereby favoring interactions with the cationic heavy 180 metal ions (Blanco A, 2000; Choi and Yun 2006, Remacle 1990). The fungus is considered to be the most efficient biosorbent organism, as its cell wall is composed of around 90% of 181 polysaccharide. Aspergillus sp. and Penicillium sp. have been reported to be most efficient in 182 183 biosorption of different metals and have been utilized in numerous sorption studies. A comparative study of Cr adsorption in bacteria, yeast, and fungi, revealed that the fungal species 184

under study (A. niger) was most efficient compared to *Streptococcus equisimilis* and *Saccharomyces cerevisiae* in adsorbing Cr (Goyal et al., 2003). Studies suggest that microbial species may be specific towards metal ions because of the variation in cellular compositions of microbial systems and the functional sites responsible for metal binding (Das et al. 2008).

Adsorption of Cd by the fungus Penicillium chrysogenum XJ-1 requires the involvement of -189 C=O and -OH groups (Xu et al. 2012). However, the amino groups of chitosan are the primary 190 191 sites responsible for heavy metal adsorption in the fungus *Penicillium chrysogenum*; and pretreatment with alkaline solution enhances adsorption efficiency (Tan and Chen 2003). In 192 subsequent studies, they subjected the fungal mycelium to amine modification by treating with 193 194 epichlorohydrin, a by-product of penicillin fermentation, and found that Ni adsorption got enhanced from 60 mg/g to 260 mg/g. Further, the modified biomass could retain the adsorption 195 efficiency to 10 cycles after treatment with dilute HCl (Tan et al. 2004). While treatment with 196 NaOH markedly improves binding efficiency, as alkaline treatment leads to the removal of 197 proteins, thereby providing more functional sites for metal binding (Baik et al. 2002). 198

199 Physicochemical parameters as ionic potential, ionic radius, ionic stability of metal ion play significant role in metal adsorption (Tsekova and Ilieva 2001). The fungus, A. niger 405 was 200 reported to efficiently adsorb Cr, Cu, Ni and Zn in the pH range of 4 to 6 and stated that at pH 201 202 value above 6, chemical precipitation inhibited the uptake of metal ions, while at pH below 3, 203 competitive effects between metal cations and hydronium ions hinder metal uptake (Filipovic-Kovacevic et al. 2000). Generally, a decrease in biosorption capacity is observed with increasing 204 metal concentration due to the saturation of metal-binding sites (Rao and Bhargavi 2013). 205 206 Studies also suggest two-fold removal of Cu in trained/adapted Talaromyces helices fungal cells 207 compared to their untrained counterparts, which were because the efficiency of Cu adsorption of mycelium was transferred to the spores (Romero et al. 2006). The whole mycelia of the fungus *A. niger, Rhizopus oryzae*, and *Mucor rouxii* were more efficient in adsorbing metals compared
to the cell materials (chitosan, glucan, and mannan) extracted from their cell walls. Of the
extracted cell materials, chitosan showed the highest metal uptake capacity in comparison to
mannan and glucan. Chitosan from *A. niger, Rhizopus oryzae*, and *Mucor rouxii* showed Cu
binding efficiency of 0.72, 0.85, and 1.13 mmol/g, respectively (Baik et al. 2002).

214 Similarly, live, and heat-inactivated biomass of fungus, Trametes versicolor immobilized in carboxymethyl cellulose was observed to adsorb Cu, Pd, and Zn efficiently. Higher metal 215 adsorption efficiency was observed in inactivated forms compared to corresponding live strains. 216 217 It was inferred that an increase in surface area due to the rupturing of a cell during cell death caused higher metal adsorption by dead biomass than live forms (Bayramoglu et al. 2003). The 218 studies that focused on deducing the mechanism behind biosorption revealed involvement of ion 219 220 exchange reaction between the negatively charged functional groups of the fungal cell wall and metal ions (Sintuprapa et al. 2000, Fan et al. 2008). Sintuprapa et al. (2000) reported the 221 222 involvement of a biphasic mode of Zn biosorption in the fungus *Penicillium*, involving the binding of Zn ions in an energy-dependent step followed by slow intracellular accumulation. 223 Ramasamy et al. (2011) speculated the involvement of a protein molecule (66 kDa) in 224 225 biosorption of Pb by the fungus Aspergillus fumigates. They utilized this sorption capacity in the mycoremediation of Pb from electronic-wastes and observed a maximum Pb removal of 85.25 226 227 %. The process was temperature-dependent as well as energy-dependent. A similar study aimed at removing metals from municipal sewage samples showed that *Sarcinnella* sp. had the most 228 potential in Pb absorption (0.32mg/l). In contrast, Aspergillus flavus, Fusarium sp., and 229 *Cladosporium* sp. showed efficient Cr adsorption (Chandrakar et al. 2014). Similarly, M. rouxii 230

IM-80 isolated from mine wastes were found to be the most promising strains for biosorption of Hg (Martinez-Juarez, 2012). Table 3 summarizes the different fungal species capable of biosorption and their biosorption capacity. Fungal functional group make-up has an integral role in metal biosorption and so their appropriate modification, induction of charge alteration can considerably enhance metal tolerance and be used for effective environment remediation application.

4.2. Bioaccumulation and compartmentalization

238 Though biosorption is the basic strategy employed to resist metal stress, yet microbes employ 239 additional strategy to fight metal toxicity, lest metal ions are transported across the extracellular membrane. In bioaccumulation, the metal ions are precipitated in different cellular organelles 240 (compartmentalization), thereby generating the non-toxic form of metal ions. Compared to 241 biosorption, bioaccumulation is feasible only in live cells and is thus a metabolism dependent 242 process (Ahalya et al. 2003). Some studies suggest that biosorption is followed by 243 244 bioaccumulation, where its accumulation follows the non-metabolic uptake of metal ions in different cellular organelles. For instance, Ezzouhri et al. (2010) reported that Penicillium sp. 245 makes use of both biosorption and bioaccumulation to resist Pb toxicity, where biosorption is 246 followed by both active and passive uptake of Pb. Transmission electron micrographs (TEM) 247 showed the presence of electron-dense bodies formed by Pb's precipitation indicating its 248 249 diffusion into the cell. Similarly, the Cd-tolerant fungus *Paxillus involutus* relies on a two-step detoxification mechanism involving the uptake and accumulation of Cd in the vacuolar 250 compartment. Cd uptake is assisted by a Ca²⁺ionophore A23187, a metabolically mediated 251 252 process dependent on potential gradient (Blaudez and Gabriel 2000). In a similar study, Cd uptake in Paxillus involutus was reported to occur in a sulfur dependent mode, where Cd 253

accumulation occurs in the form of electron-dense bodies in the vacuoles by binding to the sulfur
components. Radiotracer flux analyses suggested the accumulation of 50%, 30%, and 20% of Cd
in the cell wall, cytoplasm, and vacuole of the fungus, respectively (Otto et al. 2002). Sintuprapa
et al. 2000 suggested that bioaccumulation of Zn by *Penicillium* sp PT1 occurs in the form of
phosphate precipitates/ complexes.

Bioaccumulation of metals in fungal cells depends on several factors like metal concentration, 259 temperature, pH, other elements/ compounds. However, the significant variation is species-260 dependent, where the metabolic machinery plays the leading role. Metal uptake in *Calvatia* 261 excipuliformis, Hygrophorus virgineus, and Hypholoma capnoides significantly depend on 262 263 temperature. Variation in temperature alters the stability and configuration of the cell wall and even ionizes the chemical moieties, affecting metal uptake. A positive correlation was observed 264 between the metal content in the soil and the metal accumulated by the fungus, as the metal 265 accumulation capacity is greatly affected by the available metal concentration, soil pH, and 266 micronutrients (Elekes and Busuioc, 2011). Likewise, bioaccumulation of metal ions (Pb, Cu, 267 268 Zn, Ni, Cr and Cd) in A. niger and Aspergillus flavus showed a decrease in accumulation percentage with an increase in metal concentration which is because of the unavailability of 269 binding sites and competition among the metal ions for binding (Thippeswamy et al. 2012). 270 Bioaccumulation of Cr and Ni in Aspergillus sp. and Micrococcus sp. was reported to 271 significantly depend on temperature (Congeevaram et al. 2007). Accumulation of Cu in 272 Acremonium pinkertoniae resulted in the observation of bluish-green colored mycelium when the 273 fungus was cultured in the presence of CuSO₄. Elemental content analysis showed the 274 accumulation of 70-280 mg/g of Cu. Formation of coordinate bonds between the Cu ions, 275 nitrogen and oxygen atoms of amide and hydroxyl groups of the polysaccharides resulted in the 276

incorporation of Cu ions into the glucan-chitin complex leading to the formation of crystalloids 277 (Zapotoczny et al. 2007). Storage and mobilization of Zn in the vacuoles were recognized as the 278 possible mechanism involved in resisting Zn stress in the fungus Saccharomyces cerevisiae 279 (Wilson et al., 2012). The transport of Zn to the vacuoles as assisted by cobalt uptake 280 protein/zinc ion trans-membrane transporter (Cot1/Zrc1), belonging to ZnT type transporters is 281 282 also reported (Devirgilis et al. (2004). The Cot1/Zrc1 transporters originate from Saccharomyces *cerevisiae* ATCC 204508 / S288c strain and function in the uptake of Zn and Co ions (Devirgilis 283 284 et al. 2004). The resistance is thus assisted by Zincosomes, which are vesicles containing Zn. 285 Certain filamentous fungi as A. niger, Serpula himantioides, and Trametes versicolor also showed the remarkable potential to accumulate As when grown on a solid medium supplemented 286 with arsenopyrite (FeAsS). Accumulation of As was observed to be in the order T. versicolor >S. 287 himantioides>A. niger (Adeyemi, 2009). In another study, accumulation of Pb by the fungus 288 Aureobasidium pullulans was directly correlated to the amount of extracellular polymeric 289 290 substance (EPS) secreted by the fungus. TEM micrographs suggested that accumulated Pb could not penetrate cell interior due to the presence of EPS, while when EPS was extracted, Pb 291 efficiently penetrated the cell interior (Suh et al. 1999). Table 4 shows the variation in the metal 292 293 bioaccumulation capacity of different fungal species.

Arbuscular mycorrhizal fungi are reported to help the associated plants to overcome in the metal toxicity in moderately polluted soils. These fungi maintain a mutualistic/symbiotic relationship, with plants and increase the hosts' tolerance to heavy metals (Aly et al., 2011). There limited studies on *in-situ* bioaccumulation efficiency of mycorrhiza species. It has been reported that Zn bioaccumulation properties of bacteria *Brevibacillus* and mycorrhiza species favoured plant growth in Zn contaminated soil (Vivas et al, 2006). Endophyte *Trichoderma atroviride* F6,

which is reportedly resistant to Cd^{2+} and Ni^{2+} ions, significantly reduces Cd^{2+} and Ni^{2+} soil 300 toxicity to mustard (Cao et al, 2008). The tree Clethra barbinervis could tolerate and grow under 301 high heavy-metal concentrations because the root fungal endophytes like *Phialocephala fortinii*, 302 Rhizodermea veluwensis, and Rhizoscyphus sp. interacted to help in growth enhancement, and 303 decrease of heavy metal uptake by the plant (Yamaji et al 2016). Arbuscular mycorrhizal fungi 304 305 belonging to phylum The fungal hyphae immobilize and accumulate toxic heavy metals, thereby preventing their translocation to plant parts and prevent damage (Verma et al, 2019, Sánchez-306 307 Castro et al 2017, Zang et al 2019). Glomeromycota survive in symbiotic association with many 308 terrestrial plants. This fungus is widely used in heavy metal detoxification because of its property to adsorb, accumulate and chelate metal ions, as well as, sequestration in vacuoles (Mishra et al, 309 2019). It is also known to promot the growth of Aster tripolium roots by accumulating high 310 concentration of Cd in the vesicles. A positive association between fungal growth, mycelia 311 312 biomass and metal tolerance is observed due to their accumulation property (Carvalho et al, 313 2006). Most heavy metal tolerant plants harbor metal resistant fungi in their roots. In plants growing around mining wastes, the most species associated with their roots include Glomus, 314 Acaulospora, Fusarium, Phoma, Cladosporium, Microdiplodia, Pisolithu, Alternaria and 315 316 Peyronellaea and Sporobolomyces (Ortega-Larrocea et al, 2010, Li et al, 2012). Further, metal 317 tolerant fungal species as Aspergillus, Penicillium and Fusarium were observed to accumulate 318 and chelate heavy metals in polluted soil, thereby decreasing heavy metal toxicity and favoring 319 plant growth. These fungal communities were observed in moderately polluted soil and were effective towards heavy metals as Ni,Cd, Cu, Pb and Cr (Lin et al, 2010). The fungal endophytes 320 321 possess suitable metal chelation or sequestration systems and can remediate metal toxicity 322 through extracellular metal sequestration and precipitation, metal binding to the fungal cell walls,

intracellular sequestration and complexation, compartmentation, and volatilization (Fomina et al,2005b).

325

326 **4.3. Metal Chelation**

327 Detoxification of metal ions or metal chelation is another strategy of heavy metal resistance. In response to heavy metal stress, fungi secrete chelating molecules that bind to metal ions, 328 resulting in alleviation of toxic effects. Thiol (-SH) containing molecules, metallothioneins, 329 homogeneous and heterogeneous proteins, peroxidases, and organic acids (citric acid, oxalic 330 331 acid) serve as efficient metal detoxifying agents (Tripathi et al. 2007). These chelating molecules 332 combine with toxic metal ions leading to the formation of complex non-toxic forms of metals that get sequestrated in different cellular organelles. Studies suggest an increase in microbial 333 334 metabolism in heavily contaminated soil. This is due to rise in energy demand of microbial functional groups associated with metal ion uptake and chelation (Markowicz et al, 2016; GIller 335 et al, 2009). The functional role of some metal chelating agents has been described in detail in 336 the following sections of this review. 337

338 4.3.1. Organic acids

Certain fungi secrete organic acids in response to metal stress, which helps in solubilizing the metal ions and form respective metal oxalates. The process of metal detoxification using acids occurs both in an extracellular and intracellular manner. Oxalic and citric acids secreted by *A. niger, Penicillium* sp, and *Rhizopus* sp. have been reported to detoxify the effect of Cu and Pb. Oxalic acid is also reported to be produced in a significantly high concentration in response to heavy metals by a mycorrhizal fungal species *Rhizopogonroseolus* (Ulla et al. 2000). The role of

oxalic acids as chelators to reduce the toxic effect of heavy metals Cd, Cu, Pb, and Zn in 345 Beauveriacaledonica was reported by Fomina et alin 2005. In a similar study, the role of organic 346 acids (acetic, oxalic, and citric acid) as strong metal chelating agents was also reported. X-ray 347 absorption spectroscopic analysis revealed that Cu ions coordinated with the carboxyl (-COO⁻) 348 groups present in oxalic acid and Pb coordinated with phosphate (-PO₄³⁻) groups. It was thus 349 350 suggested that metal chelation by oxalic acids, ligand formation, and oxalate crystals production served as means of metal resistance by the fungus, B. caledonica. An increase in organic acid 351 352 production was observed in fungus Aspergillus foetidus adapted to metal ions compared to the 353 control set (Ge et al. 2011). Nickel oxalate dihydrate crystals are formed in response to Ni toxicity by a multi-tolerant strain of A.niger. However, oxalate crystals were not formed in the 354 presence of Cd, Co, Cu, and Cr, indicating a specific defense towards Ni (Magyarosy et al. 355 2002). Utilization of oxalic acid as a means to survive metal stress in certain species of wood 356 rooting fungus, including Bjerkandera fumosa, Fomitopsis pinicola, Phlebia radiata, and 357 358 *Trametes versicolor* have also been reported (Wilkolazka and Gadd 2003).

359 **4**.

4.3.2. Metal chelating compounds

360 Proteins, peptides, enzymes, and some thiol-containing molecules also assist in detoxification of metal ions. Proline, malondialdehyde, and catalase enzymes are reported to be synthesized in 361 362 response to Pb (Thippeswamy et al. 2014). Resistance to Zn in the fungus Russulaatropurpurea PRM 858109 is reported to be assisted by cysteine-containing peptides RaZBP1 and RaZBP2, 363 wherein Zn coordinates with the cysteine (Cys) and histidine (His) residues of RaZBP, helping in 364 the sequestration of excessive Zn in cellular organelles (Leonhardt et al. 2014). Similarly, 365 Aspergillus sp.P37 strain uses reduced glutathione (thiol group compounds) like compounds to 366 resist arsenate toxicity. The molecule complexes with reduced arsenate, thereby forming 367

 $As(GS)_3$, which accumulates in the vacuoles. The acidic pH in the vacuole helps to stabilize the 368 entrapped As(GS)₃ (Canovas et al. 2004). Thiol containing GSH is also known to involve the 369 370 non-enzymatic detoxification of H_2O_2 and scavenges O_2^- radicals formed in response to Cd stress in the fungus Paxillus involutus (Otto et al. 2002). However, it was also reported that at a high 371 Cd concentration, GSH was found to be ineffective because Cd binds to glutathione or causes its 372 373 diminution (Jacob et al. 2001). Cu resistance in wood-rotting fungus Phanerochaete chrysosporium, Schizophyllum commune, Daedalea quercina, and Pleurotus ostreatus was 374 375 deduced to be due to the synthesis of Cu bio-ligands. The Cu-binding ligands were of 20-60 kDa 376 size in the case of P. chrysosporium, P. ostreatus, and S. commune, while in the case of D. quercina was of a small molecular weight. Phanerochaete chrysosporium cells grown in the 377 presence of Cu showed variation in the sequence of Cu binding ligand compared to control cells 378 (Vacchina et al. 2002). 379

Melanin is another group of fungal chelators produced in response to heavy metals. They are 380 381 composed of phenolic units, peptides, carbohydrates, fatty acids, aliphatic hydrocarbons and 382 hence efficiently bind to metal ions and the metal complex accumulate in the form of electron-383 dense granules (Forgarty and Tobin, 1996). In a study conducted by Gadd and Rome (1988), melanin obtained from the fungus Aureobasidium pullulans and Cladosporium resinae showed 384 385 higher sorption for Cu ions compared to whole biomass. Pigmented biomass showed greater biosorption in comparison to albino biomass. Even the addition of melanin to the culture of 386 albino Aureobasidium pullulans in the presence of Cu showed a decrease in toxicity level. 387

Metallothiones represent the class of metal chelators composed of thiol groups and are produced in fungi, algae, and plants in response to metal stress. The involvement of metallothionein in resisting Cu and Zn in the fungus *Pisolithus tinctorius* was reported for the first time by Morselt

et al. (1986). Later, the involvement of thiol compounds (glutathione, phytochelatins, or 391 metallothioneins) in Cd detoxification was reported in the ecto-mycorrhizal fungi, Paxillus 392 393 *involutus.* An increase in glutathione and γ -glutamylcysteine content was observed on Cd treatment (Courbot et al. 2004). Recently, the possible involvement of metallothionein related 394 compounds in detoxification of Cu and Zn was reported. They suggested an increase in protein-395 396 bound disulfide and metal-thiolate clusters with increasing metal concentration. They also speculated the involvement of a cysteine-containing peptide RaZBP, sharing 77% similarity to 397 metallothionein (MT) that helps in bio-accumulating Zn similar to Zn-MT complexes for 398 399 resisting the toxic effect (Leonhardt et al. 2014).

400 **4.4. Efflux transport for metal exclusion**

401 The efflux system is quite necessary to regulate the concentration of metals in the cell interior. Different non-specific transport systems are present in microorganisms that take up both 402 essential and non-essential metals. However, when metal ions are in a concentration that proves 403 to be potentially toxic to microorganisms, specific active efflux systems assist in the exclusion of 404 405 the metals. Active transport or efflux mechanism is known to be significantly involved in conferring metal resistance to microorganisms. The metal ions accumulated in the cytoplasm are 406 excreted out from the cell interior. Bacteria mostly use this mechanism though some arsenic 407 408 resistant fungus has been reported to use this as a defense mechanism. Aspergillus P37 utilizes 409 the mechanism of extracellular efflux in addition to intracellular accumulation of arsenite to fight 410 arsenate toxicity. The fungus reduces As(V) to As(III), followed by an efflux of reduced arsenate out of the cell(Canovas et al. 2004). 411

A similar mechanism was proposed in a wild type arsenic tolerant fungus *A. niger* (Mukherjee et
al. 2010). A comparative analysis between an arsenate-resistant fungus *Hymenoscyphus ericae*

(isolated from contaminated mine sites) and non-resistant *H. ericae*, suggested the involvement 414 of a rapid efflux transport system in the resistant strain compared to non-resistant species 415 416 (Sharples et al. 2000). Arsenic uptake in both the cases occurred by passive diffusion and followed similar uptake kinetics; the arsenate resistant variety lost 90% of the absorbed arsenate. 417 In comparison, non-resistant strain showed only 40% removal. Rapid efflux of H₃AsO₃ ensues 418 419 after a reduction step wherein H_2AsO_4 is reduced to H_3AsO_3 after uptake. This mechanism of As tolerance is similar to that used by the bacteria *Staphylococcus aureus*, which reduces H₂AsO₄⁻ 420 421 to H₃AsO₃before it is removed by cells (Broer et al. 1993).

422 **5. Fungus mediated metal nanoparticle synthesis**

Fungi have the potential to reduce metal into corresponding metal ions when they adhere to or 423 424 are absorbed into the cell interior. Fungal proteins, enzymes, cofactors, and other metabolites play crucial roles in the organism's survival and reduce metal ions to their nanoparticulate forms 425 (Mehra and Winge 1991). Nanoparticles are typically in the range of 1-100 nm comprising of, 426 427 but not limited to, metals such as silver (Ag), gold (Au), iron (Fe), zinc (Zn), cadmium (Cd), copper (Cu), barium (Ba), platinum (Pt), palladium (Pd), titanium (Ti), selenium (Sn), zirconium 428 (Zr), and cobalt (Co). Such nanoparticles might be of pure metallic forms or metal compounds 429 such as oxides and sulfides (Sidiqqi et al. 2016). A common approach of synthesizing metal 430 nanoparticles is by reduction of metal ions using a suitable reducing agent. Various organic and 431 432 inorganic chemical compounds may serve as a reducing agent. Recently the use of biological entities, such as bacteria, virus, fungi, plants, and their cellular component serve as reducing 433 agents, and provide a greener, sustainable, and effective alternative. In this section, the use of 434 435 various fungi for metal nanoparticle synthesis has been discussed. Fungi being eukaryotic species produce a diverse abundance of bioactive metabolites. The greater the production of such 436

metabolites, the greater the potential for metal reduction. Fungi belonging to different classes, particularly Ascomycota and Basidiomycota, have been reported to be capable of synthesis of various metal and metal compound nanoparticles (Ovais et al. 2018; Dhillon et al. 2011; Gade et al. 2010). Fungi secrete substantial amounts of enzymes and proteins that help obtain a higher nanoparticle yield. It also offers the advantages of secure downstream processing and handling ease suitable for biotechnological processes. A general protocol followed for the fungal synthesis of metal nanoparticles is shown in Figure 3.

The studies provide insight to heavy metal tolerance and uptake by fungus which greatly exceeds the accepted permissible limit in contaminated environment. The specific metal tolerance mechanism can be adapted in clean-up of polluted soil, waste management and at mining sites for extraction and removal of metals without need of modification steps.

448

449 5.1 Mechanism of nanoparticle synthesis

Biosynthesis of metal nanoparticles using fungal cells follows either of the two mechanisms, 1) 450 intracellular or 2) extracellular synthesis routes. In intracellular synthesis, the nanoparticles are 451 formed and localized in the cytoplasm, cell wall, or cell membrane. The nanoparticle precursor 452 (metal ions) first interacts with oppositely charged cell surface moieties where they may be 453 454 simultaneously reduced to respective nanoparticles and remain bound to the cell surface. Such nanoparticles may diffuse to the cell membrane or cytoplasm. Alternatively, the ions may be 455 internalized by active or passive transport inside the cell and reduced by intracellular reducing 456 457 agents (Dhillon et al. 2011).

One of the earliest studies with an ascomycete, *Verticillium* sp. showed gold nanoparticles of 458 about 20 nm diameter formed and found localized on the walls and mostly on the inner boundary 459 of the cytoplasmic membranes of the cells (Mukherjee et al. 2001a). The intracellular synthesis 460 observed visually by the change in color of the fungal biomass to pink-purple was due to the 461 accumulation of nanoparticles within the cells. Identical synthesis reactions with different 462 463 genera of fungi ruled out the non-specific mechanism involving cell-wall sugars. The role of electrostatic interaction between the negatively charged tetrachloroaurate (AuCl₄) ions and 464 positively charged groups in the cell wall enzymes was speculated to be necessary for the 465 internalization of the ions and subsequent reduction. The nanoparticles could then diffuse inside 466 the cell and localise on the inner surface of the membrane or in the cytoplasm. The similar 467 intracellular synthesis was studied using the same species of fungus to reduce silver nitrate 468 (AgNO₃) to Ag nanoparticles of 25 ± 12 nm in diameter (Mukherjee et al. 2001b). 469

During extracellular synthesis, nanoparticles are formed outside the fungal cells. Ascomycete, 470 471 Fusarium oxysporum, was used for extracellular synthesis of Au nanoparticles from AuCl₄ ions 472 by Mukherjee et al. (2002). Enzymes released by the cells into the medium were speculated to 473 cause the reduction. The extracellular synthesis was observed by the color change of the medium from colorless to pink-purple and confirmed by TEM analysis of the solution. The study reported 474 475 a species-specific synthesis of Au nanoparticles where NADH-dependent reductases specific to F. oxysporum were believed to be facilitating nanoparticle formation. However, no specific 476 protein was identified in the study. To understand the mechanism of extracellular synthesis, 477 Duran et al. (2005) synthesized Ag nanoparticles with both mycelial biomass, as well as, cell-478 free extracts of various Fusarium strains. The reduction of the metal ions and subsequent 479 synthesis of nanoparticles took place with nitrate-dependent reductases and electron shuttle 480

quinones. The reductases, which were responsible for the reduction of Ag^+ ions to Ag^0 481 nanoparticles, were found to be species-specific. Under similar conditions, different F. 482 oxysporum strains produced extracellular Ag nanoparticles, whereas F. moniliforme did not 483 produce nanoparticles either extracellular or intracellular (Duran et al. 2005). Further, the 484 specific anthraquinones and naphthoquinones which were identified to be the electron shuttles in 485 486 F.oxysporium were not detected in F. moniliforme. It is speculated that a conjugation between the quinone electron shuttle with the reductase somehow facilitated the formation of Ag 487 nanoparticles. Hulikere et al. (2019) demonstrated the role of NADPH for the synthesis of Ag 488 489 nanoparticles through the reduction in the cell-free extract of endophytic fungus *Cladosporium* cladosporiodes. The extract was dialyzed against distilled water for 48 h. AgNO₃ solution was 490 added to the dialysate in the presence and absence of NADPH. The formation of Ag 491 nanoparticles only in the presence of NADPH was evidential of the role of an NADPH-492 dependent reductase enzyme in the formation of Ag nanoparticles. In a thermophilic fungus, the 493 494 size of the reducing agents for Au nanoparticle synthesis was found to be less than 3 kDa (Molnar et al. 2018). This could indicate that smaller molecular species such as amino acid, 495 cofactors, glucose could also serve as reducing agents. Figure 4 shows a schematic 496 497 representation of the cellular processes involved in the intracellular and extracellular synthesis of metal nanoparticles. 498

499 5.1.1. Ligand capping agents

Ligand capping agents, otherwise called stabilizing or functionalizing agents, are molecules that bind to nanoparticle surface to confer stability and prevent nanoparticle aggregation (Sharma et al. 2019). Unlike physicochemical methods of nanoparticle synthesis, which require additional steps for surface functionalization, in biogenic methods synthesis and capping occur

simultaneously (Singh et al. 2016). Likewise, in fungi, enzymes, peptides, and proteins play a 504 vital role as reducing and capping agents (Zhao et al. 2018). However, there are few reports on 505 the exact characterization and identification of the capping biomolecules in biogenic 506 nanoparticles. Some studies on the fungal synthesis of nanoparticles suggest electrostatic 507 interaction of free amine groups or cysteine residues with nanoparticle surface (Mukherjee et al. 508 509 2008, Sanghi et al. 2009). Using elemental spectroscopy imaging Gade et al. (2008) confirmed the presence of sulfur atom around Ag nanoparticles synthesized using A. niger. In this case, 510 native protein molecules stabilized the nanoparticle by binding to their surface through sulfur 511 512 atoms in their constituent amino acids.

513 Studies on fungal mediated Au nanoparticles synthesis suggest protein capping via their respective amide groups (Das et al. 2009, Binupriya et al. 2010), -CH or -OH groups (Mishra et 514 al. 2011), and aromatic and aliphatic -CN groups (Bhambure et al. 2009, Sanghi et al. 2011). 515 Metal oxide nanoparticles by fungi have also been reported to follow nitrate-reductase mediated 516 synthesis followed by protein capping (Vijayanandan et al. 2020, Abdelhakim et al. 2020). 517 Kadam et al. (2019) identified capping proteins of molecular weight 52 kDa and 58 kDa on 518 fungal ZnO nanoparticles. Likewise, Bharde et al. (2006) reported two capping proteins of 55 519 kDa and 13 kDa on fungi synthesized magnetite nanoparticles. It is evident from most of the 520 521 studies that nitrate-reductases play a central role in fungal mediated nanoparticle synthesis. 522 However, more in-depth investigations are required to identify capping biomolecules. Since their surface functionalization determines most properties of nanoparticles, specific knowledge about 523 capping proteins can help develop the targeted application of fungi-mediated nanoparticles. 524

525 **5.2. Ag nanoparticles synthesis**

Among the various nanoparticles synthesized using fungi, noble metal nanoparticles, especially Ag nanoparticles, dominate the literature. Several species of fungi reported for silver nanoparticles synthesis, along with their applications, have been enlisted in table 5.

Verticillium mediated synthesis was one of the first reports on the fungal based intracellular 529 synthesis of Ag nanoparticles (Mukherjee et al. 2001b). Mycelial biomass exposed to the AgNO₃ 530 solution produced Ag nanoparticles of about 25 nm diameter. Nanoparticles were observed to 531 532 have localized both on the cell wall and cytoplasmic membrane. Conversely, *Fusarium* based synthesis was reported to undergo an extracellular mechanism of synthesis, forming 533 comparatively smaller silver nanoparticles of 5-15 nm (Ahmad et al. 2003). The reducing agent, 534 an NADH-dependent reductase, was found specific to the species Fusarium oxysporum, and no 535 reduction or nanoparticle formation was observed with F. moniliforme. Since then, Ag 536 nanoparticles of various sizes and shapes have been reported. Extracellular synthesis of spherical 537 silver nanoparticles of about 16 nm diameter was reported using *Fusariumsolani*extract (Ingle et 538 al. 2008). Both fungi Verticillium and Fusarium oxysporum have been used to synthesize silver 539 nanoparticles, where the particles are entrapped in the biomass in the form of a film (Senapati et 540 al. 2004). Extracts of A. fumigates (Bhainsa et al. 2006), and A. niger (Gade et al. 2008) have 541 been reported to form Ag nanoparticles of 5-25 nm and 20 nm, respectively. Mycelial biomass 542 543 of A. flavus followed intracellular synthesis forming reasonably monodisperse silver nanoparticles of about 9 nm (Vigneshwaran et al. 2007). In contrast, those of A. clavatus formed 544 particles of size ranging 10-60 nm and varied morphological features, including nanohexagons 545 (Verma et al. 2010). Similarly, cell-free filtrates of *Penicillium fellutanum*, *P. brevicompactum*, 546 P. citrinum, and P. polonicum have been reported for extracellular Ag nanoparticles synthesis 547

ranging in sizes from 5 to 60 nm (Kathiseran et al. 2009; Shaligram et al. 2009; Honary et al.
2013; Neethu et al. 2018).

550 Several physicochemical factors affect the synthesis of Ag nanoparticles by fungi. pH and temperature of the synthesis medium, concentration of Ag ion precursors, and composition of 551 culture media. Various species of ascomycete Trichoderma have also been shown to 552 biosynthesize Ag nanoparticles under different conditions. The cell-free extracts of three 553 554 different strains of Trichoderma viz, T. longibrachiatum, T. viride, and T. harzianum were capable of synthesizing Ag nanoparticles in a temperature-dependent manner even within the 555 same genus of fungi (Elawami et al. 2018). In Fusarium oxysporum, it was demonstrated that 556 557 higher temperatures led to increased secretion of proteins, thereby enhancing the synthesis of Ag nanoparticles (Birla et al. 2013). The rate of nanoparticle synthesis and the size of particles are 558 also directly affected by temperature. The nanoparticle synthesis rate increased with an increase 559 in temperature, resulting in faster completion of synthesis using Aspergillus oryzae (Phanjom 560 561 and Ahmed, 2017). However, this effect was observed only in a range of temperatures, in this 562 case, between 30 to 90 °C. Abdelrahim et al. (2016), while using Rhizopus stolonifer for extracellular synthesis of Ag nanoparticles, showed that no nanoparticles were formed at both 563 564 extreme ends of their temperature study, i.e., 10 and 80 °C. This was inevitable due to the 565 inactivation and/or denaturation of biomolecules at extreme temperatures.

Another important factor considered while the synthesis of nanoparticles is the reaction pH. Using *Epicoccum nigrum*, Qian et al. (2013) established that alkaline pH favored the successful synthesis of Ag nanoparticles. Similar findings were reported by Birla et al. (2013), where nanoparticles synthesis was optimum between pH 9 and 11, which decreased with lowering pH and immediate destabilization of nanoparticles synthesized below pH 5. It is important to recall

that Ag nanoparticle synthesis by Fusariumis mediated by nitrate reductase enzymes. Hence, 571 nanoparticle synthesis under a given pH ultimately depends upon the activity of these enzymes, 572 proteins, capping agents, and activity of reducing agents under such conditions. These enzymes 573 retain their activity in the pH range of 5 to 9 in vitro (with optimum activity around pH 7) 574 (Gholami-Shabani et al. 2014). This might explain the efficient synthesis of nanoparticle in that 575 576 range. Extracellular synthesis of Ag nanoparticles was reported in the fungus Penicillium purpurogenum NPMF, which also depends on the pH of extract during nanoparticle synthesis 577 578 (Nayak et al. 2011; Pradhan et al. 2011).

The concentration of Ag ion precursors in the synthesis mixture, together with the synthesis rate, 579 580 determines the size of nanoparticles. There exists a competitive interaction between Ag ions and fungal reducing agents, which governs the morphology of the nanoparticles. The amount of 581 fungal biomass determines the concentration of reducing agents in the solution. It was shown 582 that the increase in the quantity of biomass also led to increased production of Ag nanoparticles 583 584 (Birla et al. 2013). They also presented a positive correlation between the size of nanoparticles 585 and increasing concentration of Ag ions up to a certain limit. Furthermore, excessive amounts of Ag ions led to the formation of larger particles and subsequent aggregation (AbdelRahim et al. 586 2016). 587

588 Different culture media present different growth conditions for fungi, which leads to the 589 production of different metabolites within the cells. Due to this, fungal cells cultured in different 590 culture medium have been reported to produce Ag nanoparticles with different characteristics. 591 Again Birla et al. (2013) demonstrated the effect of ten different culture media on Ag 592 nanoparticle synthesis by *Fusarium oxysporum*. They reported that fungal cells cultured in 593 MGYP medium produced maximum nanoparticles, whereas Czapek and Richard's broth

produced the least amount of nanoparticle. Supplementing culture media with specific 594 metabolites has also been reported to enhance nanoparticle production. In one study, Costa Silva 595 596 et al. (2017) used chitin enriched media in fungal culture to synthesize Ag nanoparticles using nematophagous fungus Duddingtonia flagrans. It was observed that fungal biomass growing in 597 chitin supplemented media secreted up to three times more chitinase enzymes. Also, filtrates 598 599 from chitin-fed biomass showed enhanced production of silver nanoparticles. Further investigation using FTIR and Raman scattering suggested that chitinases might serve as the 600 601 reducing and capping agent of the nanoparticles. In yet another study, Barbosa et al. (2019) 602 demonstrated that *Duddingtonia flagrans* grown in chitin enriched media produced nematicidal Ag nanoparticles. They attributed this property to chitinase coating of nanoparticles while 603 synthesis. 604

Ag nanoparticles produced using fungi have been reported to have applications in diverse fields. 605 Soil fungus Macrophominaphaseolina was used for the synthesis of protein capped Ag 606 607 nanoparticles possessing antibacterial properties against human and plant pathogenic multidrugresistant (MDR) strains of Escherichia coli and Agrobacterium tumefaciens, respectively 608 (Choudhury et al. 2014). M. phaseolina exudate was used for biosynthesis of antibacterial 609 Ag/AgCl nanoparticles and demonstrated its applicability as soybean seed protecting agent 610 611 (Spagnoletti et al. 2019). The Ag nanoparticles fabricated with Talaromyces purpureogenus also 612 had a strong anti-proliferation effect against human lung cancer cells A459 (Hu et al. 2019). Akther et al. (2019) demonstrated similar anticancer activity of Ag nanoparticles synthesized 613 with endophytic fungus Botryosphaeriarhodina. However, nanoparticles synthesized with 614 extracts of thermophilic fungus Humicola sp. were non-toxic to both normal, as well as, cancer 615 cells (Syed et al. 2013). The *Penicillium sp.* isolated from turmeric resulted in nanoparticles (25-616

617 30 nm, spherical) with good antimicrobial property against multi-drug resistant strains of
618 bacteria *Escherichia coli* and *Staphylococcus aureus* (Singh et al. 2014).

Basiodiomycete, Piriformospora indica was reported to produce Ag nanoparticles with 619 anticancer properties and antioxidant properties against three cancer cell lines, MCF-7, HeLa, 620 and HepG2 (Aziz et al. 2019). Filamentous fungus, Amylomyces rouxii, was reported for 621 extracellular biosynthesis of stable Ag nanoparticles with antimicrobial property (Mussarat et al. 622 2010). Ag nanoparticles have been widely reported for their antimicrobial efficiency and its 623 utilization in surface-enhanced Raman scattering (SERS) signaling (Jena et al. 2009; Kim et al. 624 2007). Various species of entomopathogenic fungi Metarhizhium have been reported for the 625 626 synthesis of Ag nanoparticles. Ag nanoparticles produced using *M. robertsii* exhibited antifungal and antibacterial activities (Rozalska et al. 2016). In contrast, those produced by M. anisopliae 627 were found to have potent bioinsecticidal effects on malaria vector Anopheles culicifacies, Culex 628 quinquefasciatus, and Aedes aegyptimosquitos (Litwin et al. 2020). 629

630 **5.3.** Au nanoparticle synthesis

As mentioned previously, Mukherjee and co-workers pioneered the fungi-mediated synthesis of 631 632 Au nanoparticles using fungi derived from Verticillium sp. (Mukherjee et al. 2001a). Using a similar approach, Fusarium oxysporum was employed for extracellular synthesis as well. The 633 particle size ranged between 8-40 nm (Mukherjee et al. 2002). A comparison between the roles 634 of different parts of the fungal system was presented by a rather interesting study by Xie et al. 635 636 (2007). Au nanoparticles of distinct morphologies were synthesized using fungal biomass, mycelia-free spent media, and fungal extract of A. niger in separate reactions. Fungal biomass 637 produced spherical and triangular Au nanoparticles localized on the hyphal surface, whereas the 638

spent medium synthesized triangular, truncated triangular, and hexagonal nanoplates with 639 smooth edges. The fungal extract, on the other hand, produced spirally stacked nanoplates of 640 three or more layers. This finding demonstrated the effects of different fungal metabolic 641 pathways on the overall biosynthesis of nanoparticles. Xie et al. (2007) also demonstrated the 642 temperature and pH dependency of Au nanoparticle synthesis. The reaction rate was found to 643 644 increase sharply with temperature. They further explained that higher temperatures led to fast reduction, which increased the number of Au seed particles. This further led to the formation of 645 646 smaller sized nanoparticles compared to the reaction carried out at lower temperatures. Furthermore, temperature also affected the shape of the nanoparticles; higher temperatures 647 decreased the average lateral size of the anisotropic nanoparticles. As mentioned in earlier 648 sections, here too, alkaline pH facilitated faster and better synthesis of Au nanoparticles. 649

Similarly, fungi belonging to Aspergillus sp. Penicillumsp, Verticillium sp, and Talaromyces 650 flavus have been reported for the synthesis of Au nanoparticles (Soni and Prakash, 2012; 651 Liangweil et al. 2011; Priyadarshini et al. 2014a, Priyadarshini et al. 2013; Mukherjee et al. 652 2001; Priyadarshini et al. 2014b). Previously, we have reported Talaromyces flavus based 653 synthesis of Au nanoparticles having floral morphology (Priyadarshini et al. 2013). For the 654 reduction of HAuCl₄, we compared the effects of three different reducing solutions prepared 655 656 from fungi: 1) extracellular filtrate of the fungus, 2) boiled extracellular filtrate, and 3) boiled 657 biomass filtrate. In the first solution, biomolecules were intact, whereas the latter two solutions had denatured biomolecules (presumably proteins). Though undergoing distinct reaction kinetics, 658 659 all three solutions produced Au nanoparticles. Particles from the extracellular filtrate possessed irregular floral shape, whereas nanoparticles from boiled biomass filtrate were polydispersed 660 with triangular, spherical, and hexagonal morphologies. We also reported Aspergillus candidus 661

mediated extracellular synthesis of Au nanoparticles and evaluated its applicability in visual 662 detection of cerium (Priyadarshini et al. 2015). It is speculated that the presence of negatively 663 664 charged moieties on the Au nanoparticle, rendered by the functionalization while synthesis, provided electrostatic repulsion and stability to the nanoparticles. In the presence of cerium ion, 665 neutralization of the surface charge led to the aggregation of the nanoparticles. Shift in the SPR 666 667 peak of the nanoparticle solution before and after interaction with cerium is used as basis of selective optical detection of the cerium. In yet another study, we were able to synthesize Au 668 669 nanoparticles from extracts of Aspergillus terreus and reported its antibacterial activity against 670 gram-negative bacteria (Priyadarshiniet al. 2014).

671 Aspergillus sydowii demonstrated the effect of different concentrations of fungal biomass on Au nanoparticle size (Vala 2015). Experimental setups such as shaking or stationary conditions were 672 shown to affect the morphology and localization of the Au nanoparticles (Ahmad et al. 2005). 673 674 Trichothecium sp. biomass produced spherical, rod-shaped, and triangular Au nanoparticles 675 extracellularly under stationary conditions. In contrast, it produced spherical nanoparticles 676 localized in the plasma membrane and cell wall when the reaction was carried out under shaking conditions. Other prominent ascomycetes used in Au nanoparticle synthesis include 677 Yarrowialipolytica (Agnihotri et al. 2009), Penicillium bervicompactum (Mishra et al. 2011) and 678 679 Candida ablicans (Chauhan et al. 2011). Mishra et al. (2011) demonstrated the cytotoxic effects 680 of Au nanoparticles against mouse myoblast cancer cells. Also, Chauhan et al. (2011) further conjugated Au nanoparticles with cell-specific antibodies for possible detection in liver cancer. 681 Extracts of a basidiomycete, Sclerotium rolfsii was reported to synthesize isotropic and 682 anisotropic Au nanoparticles with an average size of 25 nm (Narayanan and Sakthivel 2011). 683

684 **5.4.** Synthesis of other metal/metal compound nanoparticles

Apart from Ag and Au nanoparticles, fungal mediated nanoparticle synthesis of Cd, Cu, Pt, and 685 Se compounds, including metal oxides, and sulphides have also been reported. Fusarium seems 686 to be the most exploited fungus in the synthesis of Ag, Au, Cu, Pt, Se derived nanoparticles. 687 Ahmad et al. (2002) conducted the extracellular synthesis of CdS nanoparticles with sulfate-688 reducing enzymatic machinery of Fusarium oxysporium. The same fungus was used to 689 690 synthesize ferroelectric BaTiO₃ nanoparticles of 10 nm under ambient conditions (Bansal et al. 2006a). In yet another novel study by Bansal et al. (2006b), amorphous silica ($-SiO_2$) in rice husk 691 692 was transformed into its nanocrystalline form using F. oxysporium. Proteins present in the 693 fungus were reported to have capped and stabilized the SiO₂ nanoparticles, which then leached out into the reaction solution. Other nanoparticles successfully biosynthesized using Fusarium 694 include cadmium selenide (CdSe) quantum dots (Kumar et al. 2007), Zn nanoparticles (Bansal et 695 al. 2004) and hematite, Fe_2O_3 (Mahanty et al. 2019, Bharde et al. 2006). Mahanty et al. (2019) 696 697 showed the variation between Feoxide nanoparticles formed by three different fungal species viz. 698 Trichoderma asperellum, Phialemoniopsis ocularis and Fusarium incarnatum. Similarly, the formation of magnetite (Fe₃O₄) nanoparticles using F. oxysporium and Verticillium sp.was 699 demonstrated (Bharde et al. 2006). The Fusarium oxysporumy et again showed the capability of 700 701 Pt nanoparticle synthesis (Syed and Ahmad 2012). The mycelia biomass, when exposed to precursor ions form spherical Pt nanoparticles of 15-30 nm. Bismuth oxide nanoparticles also 702 703 were reported to be synthesized using *Fusarium oxysporium* (Uddin et al. 2008).

Another fungus frequently reported for other metal and metal oxide nanoparticle synthesis is *Aspergillus*. TiO₂ nanoparticles were synthesized using *Aspergillus flavus* (Rajakumar et al. 2012). The nanoparticles were 55-65 nm in size and possessed antibacterial property. The biosynthesis of sulfur-bearing protein-coated spherical Co oxide nanoparticles using *A. nidulans* 708 was possible (Vijayanandan and Balakrishnan 2018). A. terreus and A. oryzae were shown to biosynthesize Se nanoparticles (Zare et al. 2013 and Mosallam et al 2018). A. terreus cell-free 709 710 filtrate produced polydisperse Se nanoparticles of the size 10-100 nm. Mosallam et al. (2018) employed a different approach of synthesis by exposing the cell-free filtrate to Se ions in the 711 presence of gamma rays to maximize nanoparticle production. The ZnO nanoparticles were 712 713 synthesized using the cell-free extract of Cochliobolus geniculatus, a Zn tolerant endophytic fungus (Kadam et al 2019). Here, two proteins of 52 kDa and 58 kDa were identified to be the 714 715 capping material on the spherical nanoparticles of 2 to 6 nm. Cu nanoparticles were synthesized 716 using *Colletotrichumgloeosporiodes* biomass and then immobilized in graphene oxide-chitosan composite (Mugesh et al. 2019). The resultant nanocomposite had antibacterial and cytotoxic 717 properties against human breast cancer cells. Fungal derived crude xylanases were utilized in 718 719 synthesizing Ag-Au alloy nanoparticles with antimicrobial, antioxidant, and anticoagulant 720 properties (Elegbede et al. 2019). Aspergillus sp. has been reported to produce lead nanoparticles 721 in a size range of 1.77 to 5.8 µm (Pavani et al. 2012). A summary of different fungal strains showing the capability of nanoparticle synthesis is Table 5. 722

There are many reports of metal-based nanoparticles synthesized using a variety of fungal species. However, very few studies give conclusive information on the biomolecules involved in metal reduction leading to nanoparticle formation and the capping agents of fungal origin responsible for the stability of nanoparticles. Further studies are required to shed light on variation and diversity of biomolecules involved in nanoparticle formation from within individual strains and across fungal species. The efficiency of fungal communities to reduce metals into nanoparticle form promises cost-effective and feasible bioremediation and waste

management strategies. Green approaches with minimal resource utilization and on-siteapplication are additional advantages compared to conventional remediation methods.

732 6. Conclusion and Future Prospects

Fungi have evolved active defense mechanisms as biosorption, bioaccumulation, metal chelation, 733 and efflux transport to provide resistance from heavy metal stress in natural environments. The 734 fungal resource has emerged as a unique proposition in bioremediation of contaminated areas, 735 metal nanoparticle synthesis, and metals extracted from ores. Novel strains of fungus are being 736 737 studied for their metal extraction and nanoparticle synthesis capability. Economically feasible 738 biotechnological strategies for bioremediation of contaminated soil and water, as well as, biomining strategies of metal oxides signify the natural power of microbes. The potential risk 739 740 associated with live fungus, and the release of toxic compounds can be avoided by using heatinactivated fungus or isolated cell fractions that show an equal efficiency of nanoparticle 741 production and biosorption capacity. This would eliminate the toxicological risks and jurisdictive 742 743 constraints associated with the use of live fungus in bioremediation and water purification processes. 744

Furthermore, selection, genetic manipulation of high metal resistant fungal strains, and a mixed consortium of fungal species may enhance tolerance and resistant capacity. This would assist in the concentrated extraction of metals from dilute solutions, improving biosorption, and nanoparticle formation feasibility. The development of technology for simultaneous detection and removal of toxic metals would assist in the bioremediation of toxic metal from contaminated sites, industrial effluents and electronic-wastes. Understanding the mechanism may enable the utilization of green, non-toxic, and efficient means of bioremediation. Further interdisciplinary

studies and scaling up strategies would favour the application of these in large scalebioremediation approaches to improve unwanted outcomes in natural environments.

754

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763 Compliance with ethical standards

This article does not contain any studies with human participants performed by any of the authors.

766 **Conflict of interest**

767 The authors declare that they have no conflict of interest.

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1373 1374 1375 1376 1377 1378 1379	<u>List of Figures</u>
1380 1381	Figure 1: Schematic illustration depicting the effect of heavy metal toxicity on fungal cell
1382	Figure 2: Cellular mechanism of heavy metal tolerance by fungal cell
1383	Figure 3: General protocol for fungal mediated nanoparticles synthesis
1384	Figure 4: Schematic representation of cellular processes involved in intracellular and
1385	extracellular synthesis of metal nanoparticles
1386 1387 1388 1389	<u>List of Tables</u> Table 1: Different cellular sites that assist in binding of heavy metals to fungus
1390	Table 2: Maximum tolerance limit of some fungal strains for different metal ions
1391	Table 3: Biosorption uptake capacity of different metal ions by fungus
1392	Table 4: Bioaccumulation of different heavy metals by fungus
1393	Table 5: Synthesis of different nanoparticles by fungus
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1396 Figure 1: Schematic illustration depicting the effect of Heavy metal toxicity on fungal cell





Figure 2: Cellular mechanism of heavy metal tolerance by fungal cell





Figure 3: General protocol for fungal mediated nanoparticles synthesis



Table 1: Different cellular sites that assist in binding of heavy metals to fungus

	Location	Functional groups involved in metal uptake	Reference
	Extracellular	OH/NH ₂ and CH-OH functional groups, e.g. <i>Aspergillus fumigates</i> .	Rao et al. 2005
		-C=O and -OH groups, e.g. <i>Penicillium chrysogenum</i> XJ-1.	Xu et al. 2015
		Amino group of chitosan, e.g. <i>Penicillium</i> chrysogenum.	Tan and Chen 2003
		Phosphate binding ligands, e.g. <i>Mucor hiemalis</i> .	Hoque and Fritscher, 2019
		Electronegative interaction with outer surface of spores, e.g. <i>Mucor hiemalis</i> .	Hoque and Fritscher, 2019
	Cytoplasm	Phosphate precipitates, e.g. <i>Penicillium</i> sp PT1.	Sintuprapaet al. 2000
		Arsenate and copper precipitates, e.g. <i>Aspergillus foetidus</i> .	Ge et al. 2011
		Metal oxalates, e.g. Beauveria caledonica.	Fominaet al. 2005
	Cell organelles	Transport to vacolues by Cot1/Zrc1 transporters, e.g. <i>Saccharomyces cerevisiae</i> .	Wilson et al., 2012
		Vesicle formation (Zincosomes), e.g. S. cerevisiae.	Devirgiliset al. 2004
	Other chelators /peptides	Cysteine-containing peptides RaZBP1 and RaZBP2, e.g. <i>Russulaatropurpurea</i> PRM 858109	Leonhardt et al.2014
		Reduced glutathione, e.g. Aspergillus sp. P37;	Canovas et al. 2004
		A.niger.	Mukherjee et al. 2010
		Thiol compounds (glutathione, phytochelatins, or metallothioneins), e.g. <i>Paxillusinvolutus</i> .	Courbotet al. 2004
		Metallothionein, e.g. Pisolithustinctorius	Morseltet al. 1986
		Dihydroxynaphthalene melanin, e.g. Gaeumannomycesgraminis.	Caesar-Tonthatet al. 1995
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Table 2: Maximum tolerance limit of some fungal strains for different metal ions.

Metal	Conc (mg/l)	Fungus	Reference
Cr	1000	Aspergillus sp.	Congeevaram et al. 2007
	800	Micrococcus sp.	Congeevaramet al. 2007
	600	Penicillium chrysogenum	Levinskaite, 2002
	600	Trichodermaviride	Levinskaite, 2002
	400	Aspergillus flavus	Iramet al. 2013
	300	Trichoderma	Zhang et al. 2020
		brevicompactum QYCD-6	
	100	Fusarium solani	Iramet al. 2013
	100	Penicillium chrysogenum	Iramet al. 2013
	100	Aspergillus niger	Thippeswamy et al. 2012
	3.9	Neurospora sp SG1	Joshi 2014
Pb	1200	Alternaria alternate	Verma et al. 2016
	1600	Trichoderma	Zhang et al. 2020
		brevicompactum QYCD-6	
	1000	Aspergillus niger	Thippeswamyet al. 2012
	1000	Aspergillus flavus	Thippeswamyet al. 2012
	600	Aspergillusniger	Iramet al. 2013
	400	Geotrichumcandidum	Falih 1998
	400	Rhizopus micosporus	Oladipo et al, 2018
	400	Fomitopsismeliae	Oladipo et al, 2018
	4.5	Neurosporasp SG1	Joshi 2014
Cu	1716	Aspergillus niger(SF-5)	Akhtar et al. 2013
	1000	Rhizopus micosporus	Oladipo et al, 2018
	1000	Fomitopsismeliae	Oladipo et al, 2018
	1000	Alternaria alternate	Verma et al. 2016
	700	Pithym sp. (SF-1)	Akhtar et al. 2013
	500	Aspergillus niger	Thippeswamy et al. 2012
	400	Geotrichumcandidum	Falih 1998
	100	Aspergillus flavus	Thippeswamy et al. 2012
	5	Neurosporasp SG1	Joshi 2014
Cd	9218	Aspergillus sp.	Akhtar et al. 2013
	5732	Curvularia sp. (GF-6)	Akhtar et al. 2013
	1600	Aspergillus sp.	Akhtar et al. 2013
	600	Talaromyces helices	Romero et al. 2006
	250	Trichoderma	Zhang et al. 2020
		brevicompactum QYCD-6	
	293	Curvularia sp. (GF-6)	Akhtar et al. 2013
Zn	1000	Aspergillus flavus	Thippeswamy et al. 2012
	450	Trichoderma	Zhang et al. 2020
		brevicompactum QYCD-6	
	250	Aspergillus niger	Thippeswamy et al. 2012
	80	Penicillium sp GK1	Joshi 2014
As	500	Rhizopus micosporus	Oladipo et al, 2018
	100	Aspergillus niger	Mukherjee et al. 2010
Ni	1000	Aspergillus flavus	Thippeswamy et al. 2012

100	Aspergillus niger	Thippeswamy et al. 2012
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Table 3: Biosorption uptake capacity of different metal ions by fungus

Metal	Organism	Biomass type	Uptake efficiency	Reference
			(mg/l)	
Cr	Aspergillus niger	Live	100.3	Goyal et al. 2003
	Aspergillus flavus	Mycelium	0.25	Chandrakaret al. 2014
	Rhizopus sp.		9.50	Ahmad et al. 2005
	Fusariumsp		0.24	Chandrakaret al. 2014
	Cladosporium sp		0.44	Chandrakaret al. 2014
	Aspergillus sp.	Dead biomass	6.20	Ahmad et al. 2005
	Rhizopus sp.		2.3	Ahmad et al. 2005
	<i>Schizophyllum commune</i> Fries		18.54	Javaidet al. 2010
Pb	Aspergillus fumigates	Live	85.65	Ramasamy et al. 2011
	Penicilliumsimplicissim	Mycelium	65.60	Fan et al. 2008
	um			
	Rhizopus arrhizus		2.64	Bahadiret al. 2007
	Mucorrouxii		7.75 mg/g	Yan and Viraraghavan, 2003
	Sarcinnellasp		0.32	Chandrakaret al. 2014
	Oedogoniumsp.	Dead biomass	145	Gupta and Rastogi, 2008
	Nostocsp		93.5	Gupta and Rastogi, 2008
	Mucor rouxii		6.34 mg/g	Yan &Viraraghavan, 2003
	Trametes versicolor	Immobilized biomass (live) Immobilized biomass (inactivated)	0.85 mmol/g 1.11 mmol/g	Bayramogluet al. 2003
Cu	Aspergillus niger 405	Live Mycelium	4.4	Filipovic-Kovacevic et al. 2000
	Aspergillus niger		20.75 -93.65 mg/g	Iram and Abrar, 2015
	Mucor rouxii		2.6	Mullen et al 1992
	P. chrysosporium		3.9	Sing and Yu 1998
	Schizophyllum commune Fries	Dead biomass	21.27	Javaidet al. 2010
	Trametes versicolor	Immobilized biomass (live) Immobilized biomass (inactivated)	1.51 mmol/g 1.84 mmol/g	Bayramogluet al. 2003
Cd	Penicillium	Live	52.50	Fan et al. 2008
	simplicissimum	Mycelium		
	Penicillium		100.41	Xu et al. 2012
	cnrysogenumXJ-1		40.5	Luc et al 2010
	Knizopus cohnii		40.5	Luo et al. 2010
	Knizopus sp.		8.21	Anmad 2005
	Mucor rouxii		35.69 mg/g	Y an and Viraraghavan, 2003

	Aspergillus sp.	Dead biomass	9.5	Ahmad 2005
	Rhizopus sp.		8.21	Ahmad 2005
	Funaliatrogii		191.6	Arica et al. 2004
	Mucor rouxii		25.22 mg/g	Yan and Viraraghavan, 2003
	Trametes versicolor	Immobilized	153	Yalcinkayaet al., 2002
	Funaliatrogii	biomass	164.8	Arica et al. 2004
Zn	Penicillium sp PT1	Live	16	Sintuprapaet al. 2000
	Penicillium	Mycelium	76.90	Fan et al. 2008
	simplicissimum			
	Mucor rouxii		11.09 mg/g	Yan and Viraraghavan, 2003
	Penicillium sp PT1	Dead biomass	10	Sintuprapaet al. 2000
	Funaliatrogii		54.0	Arica et al. 2004
	Schizophyllum commune		4.83	Javaidet al. 2010
	Fries			
	Mucor rouxii		16.62 mg/g	Yan and Viraraghavan, 2003
	Funaliatrogii	Immobilized biomass	42.1	Arica et al. 2004
	Trametes versicolor	Immobilized biomass (live) Immobilized biomass (inactivated)	1.33 mmol/g 1.67 mmol/g	Bayramogluet al. 2003
Ni	Aspergillus niger 405	Live Mycelium	2	Filipovic-Kovacevic et al. 2000
	Aspergillus niger	5	23.5 mg/g	Magyarosy et al. (2002)
	Rhizopus nigiricans		1.0	Holan and Volesky 1995
	Mucor rouxii		8.46 mg/g	Yan and Viraraghavan, 2003
	<i>Schizophyllum commune</i> Fries	Dead biomass	9.0	Javaidet al. 2010
	Mucor rouxii		8.36 mg/g	Yan and Viraraghavan, 2003
Hg	Funaliatrogii	Dead biomass	403.2	Arica et al. 2004
0	Funaliatrogii	Immobilized biomass	333.0	Arica et al. 2004
	Coprinus comatus	Live mycelium	0.78 – 6.7 mg/kg	Falandysz, 2016
	Phanerochaetechryospor ium	Heat inactivated	30 -50	Kacar et al, 2002

Metal	Organism	Bioaccumulation	Reference
	_	capacity	
Cr	Aspergillus sp.	92 %	Congeevaramet al. 2007
	Aspergillus niger	25 %	Thippeswamyet al. 2012
	Micrococcus sp	90 %	Congeevaramet al. 2007
	Trichoderma	31.83 %	Zhang et al. 2020
	brevicompactum QYCD-6		
Pb	Trichoderma	80 %	Zhang et al. 2020
	brevicompactum QYCD-6		
	Aspergillus niger	75 %	Thippeswamyet al. 2012
	Aspergillus flavus	82 %	Thippeswamyet al. 2012
	Aspergillus flavus	89 %	Bano et al. 2018
	S. halophilus	60 %	Bano et al. 2018
Cu	Acremonium pinkertoniae	70-280 mg/g	Zapotocznyet al. 2007
	Trichoderma	64 %	Zhang et al. 2020
	brevicompactum QYCD-6		
	Aspergillus niger	45 %	Thippeswamyet al. 2012
	Aspergillus flavus	34 %	Thippeswamyet al. 2012
	Calvatia excipuliformis	226.52 mg/kg	Elekes and Busuioc, 2011
Cd	Aspergillus niger	0.01-0.303 mg/g	Kumar et al. 2010
	T. brevicompactum	8.45 - 20.13 %	Zhang et al. 2020
	QYCD-6		
Zn	Aspergillus niger	3.399 - 6.783	Kumar et al. 2010
		mg/g	
	Aspergillus niger	49 %	Thippeswamyet al. 2012
	Aspergillus flavus	40 %	Thippeswamyet al. 2012
	Calvatia excipuliformis	124.30 mg/kg	Elekes and Busuioc, 2011
	T. brevicompactum	2.47 - 4.43 %	Zhang et al. 2020
	QYCD-6		
	A. flavus	84 %	Bano et al. 2018
	S. halophilus	90 %	Bano et al. 2018
Ni	Aspergillus sp.	90 %	Congeevaramet al. 2007
	Aspergillus niger	41 %	Thippeswamyet al. 2012
	Aspergillus flavus	20 %	Thippeswamyet al. 2012
	Aspergillus niger	98 %	Magyarosy et al. 2002
	Micrococcus sp	55 %	Congeevaramet al. 2007

Table 4: Bioaccumulation of different heavy metals by fungus

Table 5: Synthesis of different nanoparticles by fungus

Nanoparticle	Organism	Application	Reference
Ag	Aspergillus clavatus	Antibacterial and	Verma et al. 2010
-		fungicidal activity	
	Alternaria alternata	Enhancement of	Gajbhiye et al. 2009
		antifungal activity of	
		flucanazole	
	Amylomycesrouxii	Antimicrobial activity	Mussarat et al. 2010
	Aspergillus flavus		Vigneshwaran et al. 2007
	Aspergillus fumigatus		Bhainsa et al. 2006
	Aspergillus niger	Antibacterial activity	Gade et al. 2008
	Botryosphaeria rhodina	Anticancer activity	Akther et al. 2019
	Cladosporium		Balaji et al. 2009
	cladosporioides		
	Cladosporium	Antimicrobial and	Hulikere et al. 2019
	cladosporioides	Antioxidant activity	
	Cryphonectria sp.	Antibacterial activity	Dar et al. 2013
	Duddingtonia flagrans	Nematicidal activity	Barbosa et al. 2019
	Epicoccum nigrum	Antifungal activity	Qian et al. 2013
	Fusarium oxysporum		Ahmad et al. 2003
	Fusarium oxysporum		Ingle et al. 2008
	Fusarium scirpi	Antimicrobial activity	Rodríguez-Serrano et al. 2020
	Fusarium oxysporum		Birla et al. 2013
	Humicola sp.	Cytotoxicity of nanoparticles	Syed et al. 2013
	Macrophomina	Antibacterial activity	Choudhury et al. 2014
	phaseolina (Tassi) Goid	against MDR	
	Penicillium		Shaligram et al. 2009
	brevicompactum		
	Penicillium citrinum		Honary et al. 2013
	Penicillium fellutanum		Kathiseran et al. 2009
	Penicillium polonicum	Antibacterial activity	Neethu et al. 2018
		against MDR strains	
	Penicillium sp	Antibacterial activity against MDR strains	Singh et al. 2014
	Phaenerochaete		Vighneshwaran et al.
	chrysosporium		200
	Phoma golemrata	Antibacterial activity	Birla et al. 2008
	Piriformo sporaindica	Anticancer activity	Aziz et al. 2019
	Rhizopus stonililifer		AbdelRahim et al. 2016
	Talaromyces	Antibacterial and anti-	Hu et al. 2019
	purpureogenus	proliferation activity against lung cancer cells	
	Trichodermaatroviride	Antibacterial activity	Saravanakumar et al. 2018
	Trichoderma		Elawami et al. 2018

	longibraciatum		
	Trichoderma reesei		Vabahi et al. 2011
	Trichoderma asperellum		Mukherjee et al. 2008
	Verticillium sp.		Mukherjee et al. 2001b
	Aspergillus clavatus	Antibacterial and	Verma et al. 2010
		fungicidal activity	
	Phomopsis helianthi	Antimicrobial activity	Gond et al. 2019
	Metarhizium robertsii	Antimicrobial and	Rozalska et al. 2016
		catalytic activity	
	Alternaria alternate	Enhancement of	Gajbhiye et al. 2009
		antifungal activity of	
		flucanazole	
Ag/AgCl	Macrophomina	Antibacterial activity,	Spagnoletti et al. 2019
	phaseolina	soybean seed	
	· · · · · · · · · · · · · · · · · · ·	protection	N 11 1 0001
Au	Verticillium sp.		Mukherjee et al. 2001a
	Aspergillus candidus	Cerium detection	Priyadarshini et al.
	A	Samthagia of different	2015 Via at al. 2007
	Aspergilius niger	synthesis of different	Ale et al. 2007
		nanoparticles using	
		different parts of	
		fungal system	
	Aspergillus niger		Bhambure et al. 2009
	Aspergillus sydowii		Vala 2015
	Aspergillus terreus	Antibacterial activity	Priyadarshini et al. 2014
	Aspergillus oryzae var. viridis		Binupriya et al. 2010
	Aspergillus sp. WL-Au	Catalytic reduction of p-Nitrophenol	Shen et al. 2017
	Candida ablicans	Cell-specific conjugation for liver cancer detection	Chauhan et al. 2011
	Fusarium oxysporium		Ahmad et al. 2002
	Fusarium oxysporum		Mukherjee et al. 2002
	Mariannaea sp. HJ	Catalysis of p-	Pei et al. 2017
		Nitrophenol reduction	
	Penicillium	Cytotoxic effect	Mishra et al. 2011
	bervicompactum	against mouse	
		mayoblast cancer cells	
	Sclerotiumrolfsii		Narayanan and Sakthivel 2011
	Talaromyces flavus		Priyadarshini et al. 2013
	Trichotheciumsp		Ahmad et al. 2005
	<i>Trichoderma sp</i> . WL- Go	Catalysis in degradation of aromatic pollutants	Qu et al. 2018
	1	Policianto	

	Variation line a lastice a		A amile atminest al 2000
CdS	Fusarium orosporium		Agninouri et al. 2009
	Fusarium oxysporium	Earmoalastria programta	Annual et al. 2002
BallO ₃	Fusarium oxysporium	Ferroelectric property	Bansal et al. 2006a
S_1O_2	Fusarium oxysporium	Biotransformation of	Bansal et al. 2006b
		amorphous silica to	
		nanocrystalline form	
0.10		in rice husk	V. 1 2007
CdSe	Fusarium oxysporium		Kumar et al. 2007
ZrO_2	Fusarium oxysporium		Bansal et al. 2004
Fe_2O_3	Fusarium incarnatum		Mahanty et al. 2019
Fe ₃ O ₄	Fusarium oxysporium		Bharde et al. 2006
	Phialemoniopsis		Mahanty et al. 2019
	ocularis		
	Trichoderma		Mahanty et al. 2019
	asperellum		
Pt	Fusarium oxysporium		Syed and Ahmad 2012
	Verticillium sp.		Bharde et al. 2006
TiO ₂	Aspergillus flavus	Antibacterial activity	Rajakumar et al. 2012
CoO	Aspergillus nidulans		Vijayanandan and
			Balakrishnan 2018
CoFe ₂ O ₄	Monascus purpureus	Antioxidant, anticancer	El-Sayed et al. 2020
		and antimicrobial	
		property	
Se	Aspergillus terreus		Zare et al. 2013
	Mariannaea sp HJ		Zhang et al. 2019
	Aspergillus oryzae	Antibacterial activity	Mosallam et al 2018
ZnO	Cochliobolus		Kadam et al. 2019
	geniculatus		
	Alternaria tenuissima	Antimicrobial,	Abdelhakim et al. 2020
		anticancer, antioxidant	
		band photocatalytic	
		activity	
	Cordyceps militaris	Photocatalytic	Li et al. 2019
		degradation of	
		Methylene blue dye	
Cu	Colletotrichum	Antibacterial and	Mugesh et al. 2019
	gloeosporiodes	cytotoxic property	
		against human breast	
		cancer cells	
Bi ₂ O ₃	Fusarium oxysporium		Uddin et al. 2008
Ag-Au	A. niger	Antibacterial,	Elegbede et al. 2019
	T. longibrachiatum	Anticoagulant, and	
		Antioxidant activity	
Pb	Aspergillus sp		Pavani et al. 2012